

Modelling spread of the invasive macrophyte *Cabomba caroliniana*

MICHAEL J. JACOBS AND HUGH J. MACISAAC

Great Lakes Institute for Environmental Research, University of Windsor, Windsor, ON, Canada N9B 3P4

SUMMARY

1. Predicting spread of non-indigenous species requires an understanding of where propagules are being transported, and whether these propagules can survive in the novel habitat and successfully integrate into the recipient community. In this study, we model potential spread of invading *Cabomba caroliniana* in Ontario, Canada, using a combination of passive and active dispersal models coupled with an environmental suitability model, thereby considering the first two stages of the invasion process.
2. Measures of propagule pressure incorporated both human-mediated dispersal via trailered boats, and advective flow from invaded to non-invaded systems, while habitat suitability was forecasted by combining native and global data sets and using boosted regression trees.
3. Risk of invasion differed depending on the combination of approaches used and the time period considered. Three lakes appear to be at greatest risk owing to a combination of high boater and water movement from invaded sources, and high environmental suitability. The best predictors of lake suitability were pH, mean lake temperature and dissolved calcium concentration. Hundreds of lakes in Ontario may be suitable for establishment of *Cabomba*, highlighting the need for vector management.

Keywords: boosted regression trees, environmental niche modelling, gravity model, non-indigenous species, propagule pressure

Introduction

The invasion process may be conceptualized as a series of barriers that an introduced species must overcome before successfully invading a community. These limitations include geographic barriers that restrict or preclude the introduction of propagules, adverse ambient environmental conditions and biological constraints that may affect integration into the new community (Richardson *et al.*, 2000; Colautti, Grigorovich & MacIsaac, 2006).

A number of recent reviews have highlighted the importance of propagule pressure to the success of non-indigenous species (NIS; e.g. Lockwood, Cassey & Blackburn, 2005; Colautti *et al.*, 2006; Hayes &

Barry, 2008). Propagule pressure refers to the number of inoculation events, the number of propagules introduced per event, and the condition of introduced propagules (Williamson & Fitter, 1996; Lockwood *et al.*, 2005). Propagule pressure can be very difficult to quantify, although often it can be at least semi-quantitatively assessed (e.g. Rouget & Richardson, 2003; Drake & Lodge, 2004; Herborg *et al.*, 2007). From a management context, assessments of propagule pressure provide the first step in identification of areas vulnerable to invasion. However, introduction effort can only inform where NIS are introduced, and thus potential rather than actual distribution (Rouget & Richardson, 2003).

A complementary approach seeks to identify areas vulnerable to invasion based upon environmental suitability (e.g. Thuiller *et al.*, 2005). One increasingly popular approach by which this method is applied is through application of machine-based learning

Correspondence: Hugh J. MacIsaac, Great Lakes Institute for Environmental Research, University of Windsor, Windsor, ON, Canada N9B 3P4. E-mail: hughm@uwindsor.ca

programs, such as CLIMEX or Genetic Algorithm for Rule Set Prediction (GARP) (Peterson & Vieglais, 2001; Peterson, 2003). These methods assume that propagules are available for colonization, and seek to identify areas where environmental conditions are sufficiently favourable that establishment of the NIS may be expected. Some investigators have combined vector-based and environmental niche modelling approaches to assess the invasion risk of particular areas (e.g. Herborg *et al.*, 2007).

Given that multiple barriers must be overcome for successful establishment and spread, it is not surprising that models that combine multiple limiting factors may have greater predictive ability than those based on single factors (Kolar & Lodge, 2001, 2002; Rouget & Richardson, 2003; Herborg *et al.*, 2007; Jerde & Lewis, 2007; Theoharides & Dukes, 2007).

Once established, NIS may disperse from the original invasion site via natural and/or human mechanisms (see Vander Zanden & Olden, 2008). Human-mediated dispersal may transport individuals to distances farther or in much higher numbers from the source than they could disperse naturally. Humans also may be instrumental to secondary spread of invaders following the initial colonization event (MacIsaac *et al.*, 2004; Muirhead & MacIsaac, 2005).

Cabomba caroliniana Gray or fanwort is a submersed, perennial plant that roots in sediment of both stagnant and briskly flowing waters. The species is native to South America, although it has been introduced both deliberately and unintentionally in many parts of the world, including Kasshabog Lake in Ontario, Canada (Mackey & Swarbrick, 1997). *Cabomba* can reproduce sexually via seed production and asexually via auto-fragmentation, provided there is at least one node and an intact leaf (Mackey & Swarbrick, 1997). *Cabomba* does not appear to reproduce sexually in Kasshabog Lake (J. Noel, unpublished data). The plant has serious adverse effects on invaded waterways, thus limiting its future spread is of paramount importance.

In this study, we identified lakes vulnerable to invasion by *Cabomba* using models that combine dispersal with environmental suitability.

Methods

Potential dispersal of *C. caroliniana* was assessed first by human-mediated spread on boats or boat trailers using a gravity model, followed by the application of

a hydrology model to assess advection from invaded or likely-to-be-invaded lakes.

Dispersal of *Cabomba*

Gravity models have been utilized to predict spread of numerous aquatic NIS (see Muirhead & MacIsaac, 2005). Gravity models link invaded sources with non-invaded destinations frequented by human vectors. Currently, the Kasshabog Lake system is the only area where *C. caroliniana* occurs in Ontario. In total, *Cabomba* has invaded a private lake (South Lake) along with the river that connects it to Kasshabog Lake. We conducted a survey at both launch sites located on Kasshabog Lake (source) to measure boater movement to regional non-invaded lakes (destinations). We assume that boaters inadvertently disperse the species to other lakes via viable fouled plants on trailered boats (see Johnson, Ricciardi & Carlton, 2001). The survey was conducted from August to September 2006, and included 41 boaters who reported taking trailered boats from Kasshabog Lake to other lakes. An origin-specific version of a production-constrained gravity model equation was utilized to measure the potential human-mediated spread (Haynes & Fotheringham, 1984). For the purpose of this model, other potential sources of propagules, such as aquarium stores, were ignored (Cohen, Mirotnick & Leung, 2007). Potential risk of spread was assessed as follows (Haynes & Fotheringham, 1984):

$$T_j = \frac{AO_j w_j}{D(j)}, \quad \text{where } A = \sum \left(\frac{D(j)}{w_j} \right), \quad (1)$$

where T_j is the interaction between Kasshabog Lake and lake j , A the balancing factor to measure the relative location of Lake Kasshabog to the destinations, O_j the propulsive power of Kasshabog Lake to lake j , $D(j)$ the distance decay function applied to lake j , and w_j the destination attractiveness of lake j .

Information collected from the surveys identified lakes that interacted with Kasshabog Lake, along with the strength of those interactions, effectively measuring boater movement and frequency to other lakes (O_j). It was assumed that if a boater visited many lakes in the survey, then Kasshabog Lake was visited before travelling to another lake. Destination attractiveness was based on the product of lake area and sport-fish diversity (Minns, 1990). These two factors measure

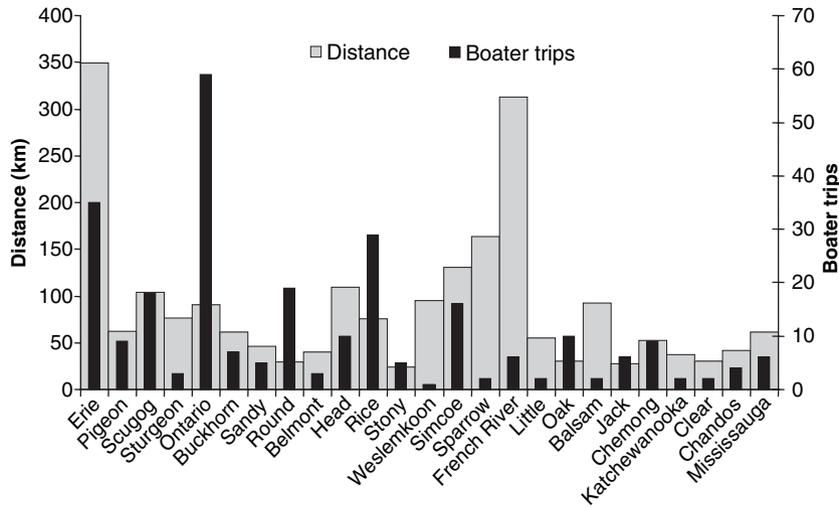


Fig. 1 The closest road distance from Kasshabog Lake (invaded source) to a destination lake (grey), along with the frequency (outflow) of those trips by trailered boats, according to surveyed data (black). Both measurements were used as variables in the origin-specific gravity model to predict human-mediated dispersal.

human attractiveness for each lake (Reed-Anderson *et al.*, 2000). Lake area was calculated by Geographic Information System (GIS) and sport-fish diversity (presence data) was obtained from the Department of Fisheries and Oceans (DFO), Burlington, ON. A distance decay function, $D(j)$, was calculated using logarithmic regression of probability of visited lakes as a function of distance from the source, separated into three intervals. All inland lakes were given an interaction value based upon eqn 1. Lakes Ontario and Erie had the greatest outflow and distances travelled from Kasshabog Lake, respectively (Fig. 1), but were not included in the model due to our focus on spread to inland lakes only.

A second model combined the gravity model and a hydrology model for those lakes identified as at risk in the former model. This model identified additional lakes at risk owing to their downstream location from lakes likely to be invaded via boater movements. The model was based on the 'water virtual flow-seamless provincial dataset' created by the Ontario Ministry of Natural Resources, and mapped using GIS. The data set is a fully connected, flow-directed, stream network with complete topological flow structure identifying connectivity and flow direction. A geometric network was created from the line layer. Next, flow directions were set on each line segment based on digitized direction. Finally, the utility network analyst extension was used to allow for downstream tracing of advective flow from a source. This procedure identified downstream lakes connected to a source.

The possibility of advective spread was explored for four lakes with the highest probability of *Cabomba* introduction from the gravity model. For downstream lakes, the distance from the closest upstream source lake was calculated along the virtual flow line layer using GIS. Water flow data was obtained from the Water Survey of Canada (Burlington, ON) for flow stations. Dispersal rate was calculated using the initial location where *Cabomba* was first reported in 1991 to where it was established by the end of 2006. An exponential cumulative distribution function was used to calculate the probability of *Cabomba* entering a non-invaded lake downstream in t years. Models were developed to project the probability of establishment after 1, 2, 5 and 10 years. Assumptions implicit to development of the hydrology model include: (i) *Cabomba* fragments move downstream at the same rate as the current (i.e. they are neutrally buoyant); (ii) flow within lakes and connecting streams is uniform; (iii) the dispersal rate (km year^{-1}) is constant; and (iv) advective flow follows the virtual dataset.

Environmental niche model

Environmental niche models were constructed to assess lakes in Ontario that provide suitable environmental characteristics for the establishment of *Cabomba*. Nine water chemistry parameters were used to develop the environmental niche model: dissolved oxygen (mg L^{-1}), dissolved calcium (mg L^{-1}), pH, mean surface water temperature ($^{\circ}\text{C}$), conductivity

($\mu\text{S cm}^{-1}$), alkalinity (mg L^{-1}), total phosphorus ($\mu\text{g L}^{-1}$), ammonia (mg L^{-1}) and nitrate ($\mu\text{g L}^{-1}$). In some cases, incomplete data were available on these parameters, although boosted regression trees (BRT; see below) allow for missing values; a minimum of four parameters were used for all lakes.

Boosted regression trees were used to develop an environmental suitability model using lake water quality parameter data (Elith *et al.*, 2006). BRT apply an iterative method that sums the weighted contribution of a successive chain of trees that are fit to the residuals from the previous tree (Friedman, 2001). The optimum number of iterations is reached when the residuals reach zero. The models were generated using the *gbm* package of R software with a training factor set at 0.70 assuming a Bernoulli distribution of presence/absence of *Cabomba* (Ridgeway, 2007). Model performance was evaluated based on the area under the receiver operating characteristic (ROC) curve (AUC), and was acceptable if significantly greater than 0.5 (random). Initially, we attempted to predict the global introduced distribution (minus Ontario) for the species using only environmental data from the native range. This model was developed using data for 96 lakes in Argentina, Paraguay, Uruguay and Brazil, but fit was quite poor (AUC = 0.591, 39.3% of the data explained). Consequently, the potential Ontario range was assessed using a model that incorporated data for the species' global distribution, except Ontario (17 147 lakes).

Once the BRT model was complete, lakes were classified as invaded if the estimated probability was greater than or equal to a threshold value based upon the shape of the ROC curve. A threshold was chosen based on the minimum distance from the upper left corner to the curve (maximum fit; Liu *et al.*, 2005). This method has been used successfully (Pearce & Ferrier, 2000; McPherson, Jetz & Rogers, 2004), despite a recent critique of the technique (Lobo, Jiménez-Valverde & Real, 2008). The R package verification was used to calculate the ROC curve. A 2×2 contingency table was constructed by pooling the number of lakes where *Cabomba* was predicted present/absent to actual present/absent data to explore model fit to Ontario.

A final model combined dispersal potential (gravity and hydrology model) with environmental suitability to provide a refined assessment of invasion risk after 1, 2, 5 and 10 years. Lake vulnerability to introduction

was first identified from the gravity model. Next, lakes that were considered a source were utilized in the hydrologic component. The overall introduction likelihood of each identified lake was calculated as the sum of probabilities of the gravity and hydrology models, with a maximum of 1. Probability of establishment was then determined by multiplying the probability of introduction with the probability of environmental suitability from the BRT model (Jerde & Lewis, 2007):

$$P(\text{establishment}) = (P(\text{gravity}) + P(\text{hydrology})) \times P(\text{BRT}). \quad (2)$$

Results

Survey results indicated that 23 lakes (excluding Lakes Erie and Ontario) were the recipients of recreational, trailered boats departing from Kasshabog Lake (Table 1). Most destination lakes were <100 km (road distance) from Kasshabog Lake (Fig. 1). A logarithmic distance decay function was calculated as $D(j) = -0.61 \ln(j) + 0.69$ ($r^2 = 0.99$). Lake Simcoe had the greatest interaction score, primarily due to its very large surface area, followed by Rice, Scugog and Pigeon Lakes.

The same group of four lakes accounted for over 84% of the interactions based on gravity scores, and were subsequently used as source lakes in the hydrologic model used to identify vulnerable downstream lakes. This model identified 14 lakes, five that were previously unidentified, with probabilities of introduction by *Cabomba* ranging from 0.01 to 0.13 after 1 year, and from 0.11 to 0.78 after 10 years (Fig. 2). According to this model, Round Lake is most vulnerable to advective invasion by *Cabomba* owing to its high inflow from a source lake.

The global occurrence of *Cabomba* was strongly related to pH and temperature (58.2% of variation; Table 2), and overall model predictive ability was high (AUC = 0.838; $P < 0.05$) (Fig. 3). Lakes were classified for presence/absence with 97.2% success, with a classification threshold of 0.11 (Table 3). The vast majority of these lakes did not contain *Cabomba*, nor were they expected to. The model correctly predicted (hit rate) two of the three waterbodies that *Cabomba* has invaded in Ontario. The invaded river was incorrectly predicted as non-invaded, while 12 additional lakes were incorrectly predicted (false

Table 1 Results of the origin-specific gravity model

Lake name	Surface area (km ²)	No. of trips	Sport-fish diversity	Score (%)
Simcoe	722	16	8	56.70
Rice	92	29	8	14.40
Scugog	34	18	17	6.54
Pigeon	52	9	20	6.52
Chemong	25	9	16	2.57
Stony	35	5	16	2.35
French River	73	6	10	2.34
Buckhorn	32	7	13	2.04
Sturgeon	45	3	19	1.73
Round	6	19	12	1.10
Balsam	48	2	15	0.93
Chandos	16	4	11	0.55
Head	9	10	7	0.42
Jack	14	6	6	0.40
Mississauga	7	6	10	0.28
Belmont	8	3	15	0.27
Oak	3	10	11	0.26
Sparrow	11	2	16	0.21
Weslemkoon	20	1	12	0.15
Sandy	4	5	11	0.15
Clear	6	2	4	0.04
Katchewanooka	4	2	4	0.02
Little	<1	2	10	<0.01

The interaction score (%) between Kasshabog Lake and a destination lake measures boater movement (possible human-mediated dispersal) of *Cabomba*. Surface area (calculated by GIS) and sport-fish diversity are measures of lake attraction and trips resemble the number of boaters leaving Kasshabog Lake to another lake. The interaction score was determined by combining these variables using eqn (1).

Table 2 Relative influence of environmental variables measured from the boosted regression trees model combining the native and global datasets to predict *Cabomba* presence in Ontario

Variable	Relative influence (%)
pH	39.9
Temperature (°C)	18.3
Dissolved calcium (mg L ⁻¹)	12.7
Conductivity (µS cm ⁻¹)	9.6
Total phosphorus (µg L ⁻¹)	7.0
Dissolved oxygen (mg L ⁻¹)	6.1
Alkalinity (mg L ⁻¹)	4.0
Ammonia (mg L ⁻¹)	1.9
Nitrate (µg L ⁻¹)	0.5

positives) to be invaded. A small number of lakes and rivers in northern Ontario were identified as having high environmental suitability, while a large cluster of lakes and rivers in south-eastern Ontario have a medium-to-high probability of environmental suitability (Fig. 4).

All 28 lakes identified at risk of introduction from the combined dispersal potential model (23 gravity + 5 hydrologic) were among the 468 lakes used in the environmental niche model, thus we were able to assign each a habitat suitability probability. After combining dispersal and environmental niche models, progressively more lakes were expected to become invaded as the time scale was extended from 1 to 10 years. This combined model suggests that Rice,

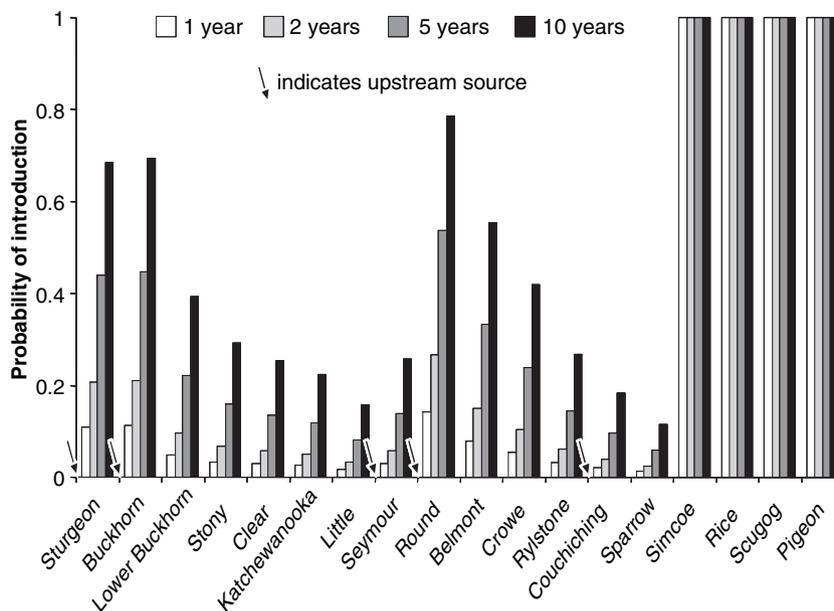


Fig. 2 Probability of *Cabomba* dispersing to a lake via advection from upstream source lakes (Simcoe, Rice, Scugog, Pigeon), identified by the gravity model, along with the current distribution of *Cabomba*, Kasshabog Lake, after 1, 2, 5 and 10 years, respectively. The probability of introduction was calculated using a cumulative exponential distribution function that used distance from an invaded, source lake to a destination lake along with a dispersal rate. Arrows indicate upstream source lakes that were assigned probabilities of 1.

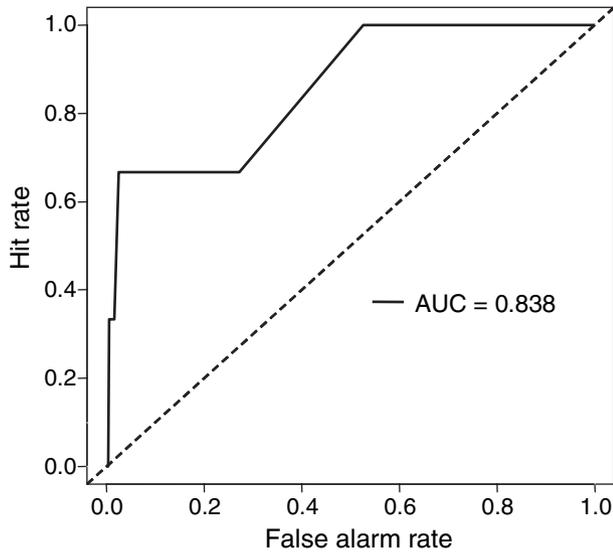


Fig. 3 Receiver operating characteristic plot of the *Cabomba* suitability model at various probability thresholds along the curve. Area under the curve is 0.838 and illustrates that the model has high predictive ability. The dashed line indicates a model that would predict no better than random.

Table 3 Contingency analysis classifying Ontario lakes predicted to have suitable habitats for *Cabomba* establishment

Predicted	Observed	
	Absent	Present
Absent	453	1
Present	12	2

Habitat suitability was generated by the native and global presence/absence dataset that the boosted regression tree model manipulated to predict the Ontario lakes' compatibility. The model explained 97.2% of data variation.

Scugog, Round and Crowe lakes have the greatest invasion risk in future years.

Discussion

Identifying which habitats are most vulnerable to biological invasion has preoccupied ecologists and managers for many years (Peterson, 2003; Richardson & Rejmánek, 2004; Theoharides & Dukes, 2007). Identification of relative invasion risk patterns should allow managers to prioritize management strategies to specific areas and/or to the vectors that transmit NIS to these areas (Basse & Plank, 2008). One approach that seems to warrant attention is the marriage of

vector-based and environmental suitability models, as they appear to provide refined estimates of invasion risk (Herborg *et al.*, 2007; Jerde & Lewis, 2007). In this study, we utilized this approach by combining two vector-based models with an environmental suitability model to project spread of the macrophyte *C. caroliniana*.

The gravity model, which utilized boater surveys to assess potential human-mediated transport, identified four lakes at high invasion risk. These lakes (Simcoe, Rice, Scugog and Pigeon) ranked the highest when considering lake area, sport-fish diversity, distance and boater movement from Kashiabog Lake (Table 1). Lake Simcoe was about average with respect to trailered boat movement from Kashiabog Lake, although its overall gravity score (and hence risk) was very high because of its very large surface area (722 km²). On the other hand, Round Lake had greater boater inflow and is much closer to Kashiabog Lake than Lake Simcoe, but was ranked tenth owing to its smaller surface area (6 km²) (Table 4). Our analysis of boater movement from Kashiabog Lake excluded Lakes Erie and Ontario. The current *Cabomba* distribution in the U.S.A. borders both of these lakes. Because these lakes had the highest boater inflow from Kashiabog Lake, they may be vulnerable to *Cabomba* introduction. Although we believe that the survey provides an accurate representation of overall traffic out of Kashiabog Lake, more extensive sampling might pick up additional lakes placed at risk by outbound boaters. The survey was responsible for identifying lakes at risk; as a result, misclassification of destination lakes would also influence invasion risk of those located farther downstream (i.e. underestimate true risk).

To better measure introduction effort, we applied a hydrology model to gauge the invasion risk associated with passive movement of viable *Cabomba* fragments among connected lakes (Boylen *et al.*, 2006). The combination of gravity and hydrology models recognized five additional lakes at risk of introduction. The most at-risk lakes based upon the combined dispersal model were Simcoe, Rice, Scugog, Pigeon, Round, Buckhorn and Sturgeon. The first four lakes (all source lakes) had the highest gravity model scores, while the final three had the shortest downstream distance from an invaded or source lake (Fig. 2). *Cabomba* has already invaded the North River and South Lake, both of which are downstream from Kashiabog Lake

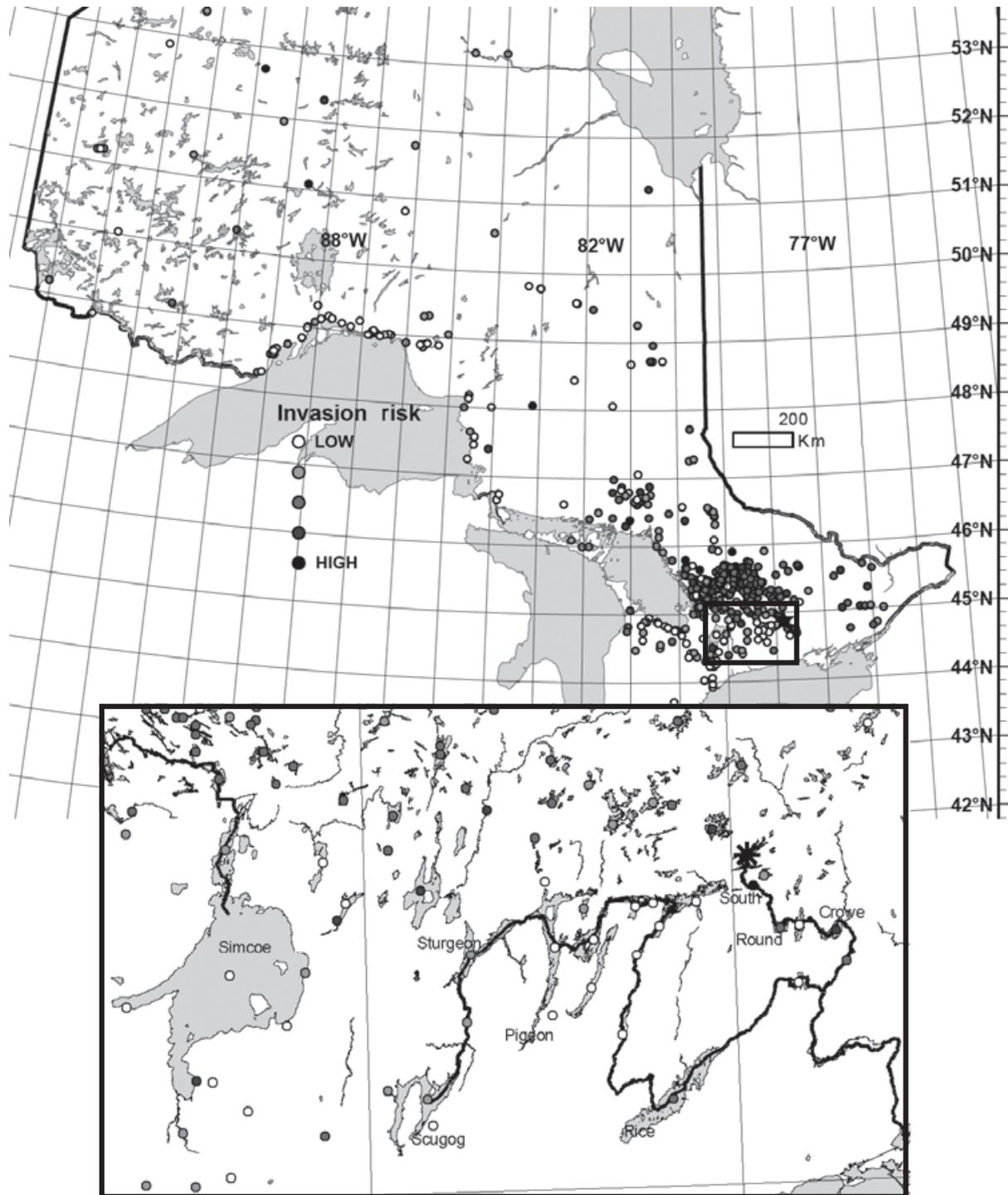


Fig. 4 A Geographic Information System representation of the ecological niche model highlighting Ontario lake suitability invasion risk based upon environmental criteria only. Inset shows the collection of lakes that are at risk due to advective dispersal from downstream connected lakes identified by the solid black line. Lakes >320 000 hectares are shown and the asterisk denotes Kasshabog Lake.

(Fig. 4 inset). The combined dispersal model differs from the advection only model in that a greater number of lakes were designated an introduction risk,

accounting for human-mediated dispersal potential. Our study considered only boater movement and advection to predict risk of *Cabomba* spread, although

Table 4 Top 10 lakes vulnerable to invasion, listed in descending order, for each of the four models

Gravity model	Gravity + hydrology model	Environmental niche model	Gravity + hydrology + environmental niche model
Simcoe	Simcoe	Tasso	Rice
Rice	Rice	Nebskwashi River	Scugog
Scugog	Scugog	Manitouwaba	Round
Pigeon	Pigeon	Alfred	Crowe
Chemong	Round	Bat	Sturgeon
Stony	Buckhorn	Missinaibi River	Pigeon
French River	Sturgeon	Three Mile	Simcoe
Buckhorn	Belmont	Moot	Buckhorn
Sturgeon	Crowe	Fawn	Rylstone
Round	Lower Buckhorn	Pipestone River	Belmont

Cohen *et al.* (2007) determined that the plant is sold extensively in pet stores, which would serve as another vector of introduction to lakes.

Our environmental niche model identified 12 lakes and rivers with high habitat suitability, suggesting that these systems are vulnerable to establishment of the species should it be introduced. If the threshold is altered (i.e. reduced) to that of the North River, which has a *Cabomba* population, then an additional 207 waterbodies were identified as suitable habitats. Thus, *Cabomba* may find numerous suitable habitats for establishment in Ontario unless introduction is prevented.

pH, temperature and dissolved calcium were identified as the best predictors of *Cabomba* presence in the boosted regression tree model (Table 2). Two rivers in Northern Ontario (51° and 52°N latitude) were predicted as suitable habitat, illustrating that *Cabomba* is not limited to tropical areas. This is consistent with the occurrence of the species in the Loosdrecht lakes, the Netherlands, which share similar latitude with Northern Ontario, even though climate differs between the areas (Schooler, Cabrera & Julien, 2008). These patterns indicate that temperature alone may not provide an accurate reflection of actual or potential occurrence of *Cabomba* (van der Heide *et al.*, 2006).

Ecological integration into the recipient community is the final consideration in the stage-specific approach to identification of invasion risk (Colautti *et al.*, 2006). Charles Elton (1958) identified a number of biological factors, notably competition and predation, which can hinder the invasion success of NIS. It is not clear whether competition is likely to impede spread of *Cabomba* in lakes with substantial introduction effort and high environmental suitability, as Capers

et al. (2007) demonstrated that the species is capable of invading systems already populated by other macrophyte species.

Forecasting invasions is an important yet imprecise science (Guisan & Thuiller, 2005). The emergence of propagule pressure and climatic suitability as key predictors of invasion success hold promise for advances in predicting vulnerability of sites to NIS invasions (Richardson & Rejmánek, 2004; Lockwood *et al.*, 2005). By combining models that embrace propagule pressure (Cohen *et al.*, 2007) with those that encompass niche-based modelling (Thuiller *et al.*, 2005), predictive power should be increased (Crossman & Bass, 2008). Other studies have coupled predictive models (Herborg *et al.*, 2007; Jerde & Lewis, 2007), although the current study is the first to utilize both active (human-mediated) and passive (advective flow) movement of propagules with analyses of environmental suitability. Our study illustrated that different combinations of lakes were deemed most vulnerable to invasion when vector-based and environmental niche models were utilized. A final model that incorporated elements of both vector- and niche-based models was most similar to the results of vector-based model that incorporated both passive and active dispersal, and highly dissimilar to the predictions from the environmental niche model. With so few lakes presently invaded in Ontario, it will take some time before the accuracy of the different models developed here can be validated.

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References

- Basse B. & Plank M. (2008) Modeling biological invasions over homogeneous and inhomogeneous landscapes using level set methods. *Biological Invasions*, **10**, 157–167.
- Boylen C.W., Eichler L.W., Bartkowski J.S. & Shaver S.M. (2006) Use of Geographic Information Systems to monitor and predict non-native aquatic plant dispersal through north-eastern North America. *Hydrobiologia*, **570**, 243–248.
- Capers R.S., Selsky R., Bugbee G.J. & White J.C. (2007) Aquatic plant community invasibility and scale-dependent patterns in native and invasive species richness. *Ecology*, **88**, 3135–3143.
- Cohen J., Mirotnick N. & Leung B. (2007) Thousands introduced annually: the aquarium pathway for non-indigenous plants to the St Lawrence Seaway. *Frontiers in Ecology and the Environment*, **5**, 528–532.
- Colautti R.I., Grigorovich I.A. & MacIsaac H.J. (2006) Propagule pressure: a null model for biological invasions. *Biological Invasions*, **8**, 1023–1037.
- Crossman N.D. & Bass D.A. (2008) Application of common predictive habitat techniques for post-border weed risk management. *Diversity and Distributions*, **14**, 213–224.
- Drake J.M. & Lodge D.M. (2004) Global hotspots of biological invasions: evaluating options for ballast-water management. *Proceedings of the Royal Society of London, Series B, Biological Sciences*, **271**, 575–580.
- Elith J., Graham C.H., Anderson R.P. *et al.* (2006) Novel methods improve prediction of species' distribution from occurrence data. *Ecography*, **29**, 129–151.
- Elton C.S. (1958) *The Ecology of Invasions by Animals and Plants*. Methuen and Company, London, U.K.
- Friedman J.H. (2001) Greedy function approximation: a gradient boosting machine. *Annals of Statistics*, **29**, 1189–1232.
- Guisan A. & Thuiller W. (2005) Predicting species distribution: offering more than simple habitat models? *Ecology Letters*, **8**, 993–1009.
- Hayes K.R. & Barry S.C. (2008) Are there any consistent predictors of invasion success? *Biological Invasions*, **10**, 483.
- Haynes K.E. & Fotheringham S. (1984) *Gravity and Spatial Interaction Models*. Sage Publications, Inc., Beverly Hills, CA, U.S.A.
- van der Heide T., Roijackers R.M.M., van Nes E.H. & Peeters E.T.H.M. (2006) A simple equation for describing the temperature dependent growth of free-floating macrophytes. *Aquatic Botany*, **84**, 171–175.
- Herborg L.-M., Jerde C.L., Lodge D.M., Ruiz G.M. & MacIsaac H.J. (2007) Predicting invasion risk using measures of introduction effort and environmental niche models. *Ecological Applications*, **17**, 663–674.
- Jerde C.L. & Lewis M.A. (2007) Waiting for invasions: a framework for the arrival of nonindigenous species. *The American Naturalist*, **170**, 1–9.
- Johnson L.E., Ricciardi A. & Carlton J.T. (2001) Overland dispersal of aquatic invasive species: a risk assessment of transient recreational boating. *Ecological Applications*, **11**, 1789–1799.
- Kolar C.S. & Lodge D.M. (2001) Progress in invasion biology: predicting invaders. *Trends in Ecology and Evolution*, **16**, 199–204.
- Kolar C.S. & Lodge D.M. (2002) Ecological predictions and risk assessment for alien fishes in North America. *Science*, **298**, 1233–1236.
- Liu C.R., Berry P.M., Dawson T.P. & Pearson R.G. (2005) Selecting thresholds of occurrence in the prediction of species distributions. *Ecography*, **28**, 385–393.
- Lobo J.M., Jiménez-Valverde A. & Real R. (2008) AUC: a misleading measure of the performance of predictive distribution models. *Global Ecology and Biogeography*, **17**, 145–151.
- Lockwood J.L., Cassey P. & Blackburn T. (2005) The role of propagule pressure in explaining species invasions. *Trends in Ecology and Evolution*, **20**, 223–228.
- MacIsaac H.J., Borbely J., Muirhead J. & Graniero P. (2004) Backcasting and forecasting biological invasions of lakes. *Ecological Applications*, **14**, 773–783.

- Mackey A.P. & Swarbrick J.T. (1997) The biology of Australian weeds 32. *Cabomba caroliniana* Gray. *Plant Protection Quarterly*, **12**, 155–165.
- McPherson J.M., Jetz W. & Rogers D.J. (2004) The effects of species' range sizes on the accuracy of distribution models: ecological phenomenon or statistical artefact? *Journal of Applied Ecology*, **41**, 811–823.
- Minns C.K. (1990) Patterns of distribution and association of freshwater fish in New Zealand. *New Zealand Journal of Marine and Freshwater Research*, **24**, 31–44.
- Muirhead J.R. & MacIsaac H.J. (2005) Development of lakes as hubs in an invasion network. *Journal of Applied Ecology*, **42**, 80–90.
- Pearce J. & Ferrier S. (2000) An evaluation of alternative algorithms for fitting species distribution models using logistic regression. *Ecological Modelling*, **128**, 127–147.
- Peterson A.T. (2003) Predicting the geography of species' invasions via ecological niche modeling. *Quarterly Review of Biology*, **78**, 419–433.
- Peterson A.T. & Vieglais D.A. (2001) Predicting species invasions using ecological niche modeling: new approaches from bioinformatics attack a pressing problem. *BioScience*, **51**, 363–371.
- Reed-Anderson T., Bennett E.M., Jorgensen B.S., Lauster G., Lewis D.B., Nowacek D., Riera J.L., Sanderson B.L. & Stedman R. (2000) Distribution of recreational boating across lakes: do landscape variables affect recreational use? *Freshwater Biology*, **43**, 439–448.
- Richardson D.M. & Rejmánek M. (2004) Conifers as invasive aliens: a global survey and predictive framework. *Diversity and Distribution*, **10**, 321–331.
- Richardson D.M., Pysek P., Rejmánek M., Barbour M.G., Panetta F.D. & West C.J. (2000) Naturalisation and invasion of alien plants: concepts and definitions. *Diversity and Distributions*, **6**, 93–107.
- Ridgeway G. (2007) *Generalized Boosted Models: A Guide to the gbm Package*. R package version 1.5–7. Available at: <http://finzi.psych.upenn.edu/R/library/gbm/html/00index.html>. (last accessed on 2 September 2008).
- Rouget M. & Richardson D.M. (2003) Inferring process from pattern in plant invasions: A semimechanistic model incorporating propagule pressure and environmental factors. *The American Naturalist*, **162**, 713–724.
- Schooler S.S., Cabrera G.C. & Julien M.H. (2008) Ecology and biological control of *Cabomba caroliniana*. In: *Weed Biological Control with Arthropods in the Tropics* (Eds R. Muniappan, G.V.P. Reddy, A. Raman & V.P. Gandhi), pp. ???–???. Cambridge University Press, Cambridge, U.K.
- Theoharides K.A. & Dukes J.S. (2007) Plant invasion across space and time: factors affecting nonindigenous species success during four stages of invasion. *New Phytologist*, **176**, 256–273.
- Thuiller W., Richardson D.M., Pysek P., Midgley G.F., Hughes G.O. & Rouget M. (2005) Niche-based modeling as a tool for predicting the risk of alien plant invasions at a global scale. *Global Change Biology*, **11**, 2234–2250.
- Vander Zanden M.J. & Olden J.D. (2008) A management framework for preventing the secondary spread of aquatic invasive species. *Canadian Journal of Fisheries and Aquatic Sciences*, **65**, 1512–1522.
- Williamson M. & Fitter A. (1996) The characters of successful invaders. *Biological Conservation*, **78**, 163–170.

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